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# Single and binary adsorption of reactive dyes from aqueous solutions onto clinoptilolite

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# ARTICLE INFO

# ABSTRACT

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Keywords: Clinoptilolite Reactive dyes Adsorption isotherms Kinetics models Single and binary adsorption The adsorption of Reactive Blue 21 (RB21) and Reactive Red 195 (RR195) onto clinoptilolite type natural zeolite (ZEC) has been investigated at 298.15 K. The uptake of single and binary reactive dyes from aqueous solutions has been determined by UV-vis spectroscopy. Two mono-component (RB21 and RR195) and binary component (RB21 with RR195, and RR195 with RB21), isotherms were determined. The mono-component Langmuir isotherm model was applied to experimental data and the isotherm constants were calculated for RB21 and RR195 dyes. The monolayer coverage capacities of clinoptilolite for RB21 and RR195 dyes in single solution system were found as 9.652 and 3.186 mg/g, respectively. Equilibrium adsorption for binary systems was analyzed by using the extended Langmuir models. The rate of kinetic processes of single and binary dye systems onto clinoptilolite was described by using two kinetics adsorption models. The pseudo-second-order model was the best choice among the kinetic models to describe the adsorption behaviour of single and binary dyes onto clinoptilolite.

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# 1. Introduction

Synthetic dyes and pigments released into the environment mainly in the form of wastewater effluents by textile, leather and printing industries cause severe ecological problems. Wastewater from the textile industry is a complex mixture of many polluting substances ranging from organochlorine-based pesticides to heavy metals associated with dyes or the dyeing process. The reactive dyes, which represent the largest class of dyes used in textile processing industries, are almost azo compounds, i.e. molecules with one or several azo bridges (N=N) linking substituted aromatic structures. These dyes are designed to be chemically and photolytically stable, they exhibit a high resistance to microbial degradation and are highly persistent in natural environment. The release of these compounds into the environment is undesirable, not only for aesthetic reasons, but also because many azo dyes and their breakdown products are toxic and/or mutagenic for life.

Various physicochemical and biological techniques can be employed to remove dyes from wastewaters. They include the membrane coagulation/flocculation [1], ion exchange [2], advanced oxidation (chlorination, ozonation) [3], decolourisation and metabolism of the commercially used Remazol dye by a strain

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of P. Chrysosporium [4], and biological treatment (bacterial and fungal biosorption), biodegradation in aerobic or anaerobic conditions [5].

In comparation with other techniques adsorption is superior in simplicity of design, initial cost, ease of operation and insensitivity to toxic substances. Adsorption is one of the most effective physical processes for colour removal. This technique uses a large number of suitable sorbents as activated carbon [6,7].

Activated carbon is normally employed for this purpose, however, due to its high price it has not enjoyed wide-scale application. Therefore, other inexpensive and effective adsorbents have been tested for example; biopolymer [8,9] and various low-cost adsorbents as clays and zeolites material, etc. [10]. Natural zeolite adsorption is used in the chemical process industries and is playing important role in cleaning up plant effluent and municipal wastewater. Especially, clinoptilolite is probably the most abundant zeolite in nature because of its wide geographic distribution and large size of deposits. The presence of 4.5 million tons of natural zeolites of high quality, mainly those of clinoptilolite in Turkey, created an impetus for the utilization of clinoptilolite in wastewater treatment.

This study was performed to investigate adsorption of single and binary reactive dyes aqueous solutions on clinoptilolite. The constants parameters obtained from single (RB21 and RR195) adsorption equilibrium data were used to predict the binary adsorption behaviour of reactive blue21 and reactive red195 [11].



Scheme 1. (a) Possible structure of the reactive blue21. R - reactive linker arms, (b) reactive red195.

The kinetics and isotherms for dyes adsorption onto clinoptilolite (ZEC) were studied in single and binary dye solutions.

#### 2. Materials and methods

# 2.1. Chemicals and materials

The clinoptilolite was obtained from the rhyolitic tuff level of the Neogene volcano sedimentary sequence in Gördes, West Anatolia, Turkey. Clinoptilolite (ZEC) was used as adsorbents for removal of RB21 and RR195 from aqueous solutions in this study. This adsorbent was used directly for adsorption experiments without any treatments. Clinoptilolite is a species of zeolite group minerals. The typical unit cell formula of natural zeolite mineral, clinoptilolite, is (Ca, Na, K)<sub>6</sub>[(AlO<sub>2</sub>)<sub>6</sub>(SiO<sub>2</sub>)<sub>30</sub>]·24H<sub>2</sub>O. It contained about 91% clinoptilolite with the chemical composition of 67.83-66.70% SiO<sub>2</sub>, 11.63-12.04% Al<sub>2</sub>O<sub>3</sub>, 2.35-2.10% CaO, 1.59-0.48% MgO, 0.95% Fe<sub>2</sub>O<sub>3</sub>, 0.28-0.29% Na<sub>2</sub>O, 1.59-4.97% K<sub>2</sub>O, % 0.05 SO<sub>3</sub>, loss on ignition 11.81% as reported by Esenli [12]. The clinoptilolite (ZEC) was sieved to obtain a particle size smaller than 125 µm. The dyes reactive blue21 (RB21) and reactive red195 (RR195) were purchased from a textile industry in the company (Turkey) region. Colour index name of Türkisblau G133 is reactive blue21 (RB21). This dye is vinyl sulfone dyes and it has chemical class copper-phtalocyanin (C32H16CuN8 molecular weight 576.10 g mol<sup>-1</sup>) and reactive linker arms of the following structure R: SO<sub>2</sub>NHC<sub>6</sub>H<sub>4</sub>SO<sub>2</sub>CH<sub>2</sub>CH<sub>2</sub>OSO<sub>3</sub>H (molecular weight of single  $R = 550 \text{ g mol}^{-1}$  [13–16]. Colour index name of Synozol Red HF-6BN 150% is reactive red195 and this dye has a chemical class monoazo and molecular weight 1136.30 g mol<sup>-1</sup>. Stock solutions of  $300 \text{ mg L}^{-1}$  were prepared with double-distilled water. The values pH of RB21 and RR195 were measured 7.60 and 6.60, respectively. The molecular structures of the two dyes are illustrated in Scheme 1.

# 2.2. Single adsorption studies

Stock solutions  $300 \text{ mg L}^{-1}$  of RB21 and RR195 were used in this study. These stock solutions were diluted to give standard solutions from 10 to  $250 \text{ mg L}^{-1}$ . 0.1 g ZEC and 10 mL single-component solution were shaken using a shaker with a water bath to control temperature at 25.0 °C. The liquid and solid phases were separated by centrifuging at 3950 rpm for 5 min. The change of adsorbate concentration in each of the solutions was determined spectrophotometrically. All the spectrophotometric measurements were made with a CHE-BIOS Optimum UV-vis spectrophotometer. The maximum absorbance value of RB21 and RR195 was measured at 660 and 540 nm, respectively as shown in Fig. 1. The calibration graph of absorbance versus concentration obeyed a linear Lambert–Beer relationship [16].

The color removal efficiency of the dye was calculated as follows

Removal efficiency (%) = 
$$\frac{(C_o - C_e)}{C_o} \times 100$$
 (1)

where  $C_o$  and  $C_e$  are the initial concentration and concentrations of the dye at equilibrium (mg L<sup>-1</sup>), respectively.

The mass of reactive dyes adsorbed per unit mass of the adsorbents (q) at any time  $(C=C_t)$  and at equilibrium  $(C=C_e)$  was calculated from the Eq. (2).

$$q = (C_o - C)\frac{V}{W} \tag{2}$$

*q* is the adsorbed amount of reactive dyes in the solid phase at equilibrium  $(q=q_e)$  and time  $t \pmod{(q=q_t)}$ , respectively. *V* indicates the volume of dye solution (L) and *W* is the weight of the adsorbent (g).

# 2.3. Binary adsorption studies

For binary adsorption system, several initial concentrations of RR195 ( $10-250 \text{ mg L}^{-1}$ ) were prepared in 8 different glassstoppered conical flasks with the presence of constant initial concentration of  $50 \text{ mg L}^{-1}$  of RB21 in each flask. Then, the experiment was carried out following the steps as for single system adsorption process. The remaining concentrations of both solutions were also analyzed using UV-vis spectrophotometer through the multi-component determination option. The experiments were repeated for other constant initial concentrations of RB21, which are 100 and  $150 \text{ mg L}^{-1}$ . The tests were repeated over again in



Fig. 1. Typical spectrum of the single dye (a) RR195 (b) RB21.

# Table 1

Initial system conditions for equilibrium adsorption isotherms and kinetics.

Adsorption system	$\text{ZEC}\left(g\right)\pm0.002$	Dye initial concentration range (in mg L <sup>-1</sup> )
RB21	0.10	RB21: 10-250
RB21 with RR195	0.10	RB21: 50
		RR195: 10-250
RB21 with RR195	0.10	RB21: 100
		RR195: 10-250
RB21 with RR195	0.10	RB21: 150
		RR195: 10-250
RR195	0.10	RR195: 10-250
RR195 with RB21	0.10	RR195: 50
		RB21: 10-250
RR195 with RB21	0.10	RR195: 100
		RB21: 10-250
RR195 with RB21	0.10	RR195: 150
		RB21: 10-250

order to study the adsorption behaviour of RB21 with the presence of different constant initial concentrations of RR195 (50, 100 and 150 mg  $L^{-1}$ ), as given in Table 1 .

# 2.4. Adsorption kinetics studies

Adsorption kinetics study was carried out in order to test the relationship between contact time and dye uptake. The influence of the initial concentration of the reagent on the adsorption kinetic at constant temperature of 25 °C was studied. The following initial concentrations of dyes were used: single dyes concentrations 50, 100, 150 mg L<sup>-1</sup>, binary dyes concentrations 50 + 50, 100 + 100, 150 + 150 mg L<sup>-1</sup>. 0.1 g ZEC and 10 mL single and binary component solution were shaken at 25.0 °C while being rinsed with a water bath at a speed of 140 rpm. The samples were subjected to centrifugation at 3950 rpm for 5 min. A slower process follows the initial fast process. Although adsorption is very fast up to 25 min, the adsorption equilibrium is established in 90 min. Adsorption data up to 45 min were used for kinetic calculations. Amount of dye adsorbed at any time was calculated from the concentration changes during adsorption process as follows Eq. (2).

# 3. Results and discussion

# 3.1. Color removal efficiency in single and mixed dye solutions

The adsorption potential of ZEC was assessed for reactive dyes and the color removal efficiencies were measured in single and mixed dye solutions. As shown in Table 2, ZEC exhibited high adsorption efficiency for RB21. Also, comparing with the results in single dye solutions, a reduction in removal efficiency of the individual dye in the mixed dye systems was observed, although the extent of reduction was different among the dyes.

# Table 2

Removal efficiency values of	of single and binary	reactive dyes systems.
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Component	$C_{o} (mg L^{-1})$	$C_e(\mathrm{mg}\mathrm{L}^{-1})$	Removal efficiency %
RB21	50 100	21.36 51.33	57 49
	150	71.46	52
RB21	50+50	27.3+42	31
with	100+100	73.0+100	14
RR195	150+150	117+150	11
	50	45	10
RR195	100	96	4.0
	150	145	3.3
RR195	50+50	40+50	10
with	100+100	95+98	3.5
RB21	150+150	140+150	3.3

The removal efficiency of RB21 was reduced from 49% (in single solution  $100 \text{ mg L}^{-1}$ ) to 14% (in binary solution of  $100 \text{ mg L}^{-1}$ RB21 +  $100 \text{ mg L}^{-1}$ RR195). But the removal of RR195 was reduced from 4.0% (in single solution  $100 \text{ mg L}^{-1}$ ) to 3.5% (in binary solution of  $100 \text{ mg L}^{-1}$ RR195 +  $100 \text{ mg L}^{-1}$ RB21). The reduction in removal of RB21 was larger than that of RR195, as shown in Table 2. When removal efficiency values compared between single RR195 and RR195 + RB21, it was observed that these values did not change for each systems of RR195. As a result, it was revealed that RB21 was preferentially removed in mixed dye systems onto ZEC.

#### 3.2. Single solute equilibrium studies

The equilibrium adsorption data of the single and binary systems were fitted using Langmuir isotherm models. The linearized equation of Langmuir model is as in shown by Eq. (3).

$$\frac{1}{q_e} = \frac{1}{q_m b C_e} + \frac{1}{q_m} \tag{3}$$

where  $q_e$  is the adsorbed dye amount per unit weight of adsorbent at equilibrium (mg g<sup>-1</sup>),  $C_e$  is the equilibrium concentration of adsorbate (mg L<sup>-1</sup>),  $q_m$  (mg g<sup>-1</sup>) and b (L mg<sup>-1</sup>) are the Langmuir constants related to the maximum adsorption capacity and the energy of adsorption, respectively. The constant value of  $q_m$  is obtained from the slope of linear plot of  $1/q_e$  versus  $1/C_e$  as shown in Fig. 2. The essential characteristics by a separation factor  $R_L$ ; which is defined by the following equation [17],

$$R_{\rm L} = \frac{1}{1 + bC_o} \tag{4}$$

where  $C_0$  is the highest initial solute concentration and b is the Langmuir's adsorption constant (Lmg<sup>-1</sup>). The value of separation factor  $R_L$ , indicates the nature of the adsorption process as given below:

R <sub>L</sub> value	Nature of adsorption process
$R_{\rm L} > 1$	Unfavourable
$R_{\rm L} = 1$	Linear
$0 < R_{\rm L} < 1$	Favourable
$R_{\rm L} = 0$	Irreversible

The Langmuir constants are given in Table 3. The calculated dimensionless separation factors, R<sub>L</sub> are 0.721 and 0.123 for RR195 and RB21, respectively. The values are less than 1 and greater than 0. The  $R_{\rm L}$  values show that favorable adsorptions of RB21 and RR195 take place on ZEC. Zeolite nanocrystals can act as hosts for supramolecular organization of molecules. Dyes have the tendency to form aggregates even at low concentration. Such aggregates are known to cause fast thermal relaxation of electronic excitation energy. The role of the zeolite is to prevent this aggregation and to superimpose a specific organization. Dye molecules of appropriate size are arranged with their long molecular axis along the one-dimensional channels and they cannot glide past each other. This allows the filling of specific parts of the nanocrystals with a desired type of dye because of the long molecular length of RB21 and RR195. RB21 can be considered to has affinity for ZEC, because the long molecular length of RB21 is longer than that of RR195. Therefore, the maximum adsorption capacity value of RB21 onto ZEC is higher than that of RR195, as seen in Table 3.

## 3.3. Extended Langmuir equation for binary systems

It is the most generally used model in multi-component system, which permitted derived from the basis of single-component Langmuir model (Eq. (3)). The extended Langmuir is presented by the



Fig. 2. Langmuir adsorption isotherm of reactive dyes on ZEC at 25 °C (a) RR195 (b) RB21.

 Table 3

 Langmuir constants for the adsorption of RB21 and RR195.

Component	$C_o(\mathrm{mg}\mathrm{L}^{-1})$	$q_m(\mathrm{mg}\mathrm{g}^{-1})$	b (L mg <sup>-1</sup> )	$R^2$	$R_{\rm L}$
RB21	10–250	9.652	0.0291	0.99	0.123
RR195	10–250	3.186	0.0015	0.99	0.721

following equation [18].

$$q_{ej} = \frac{q_j^o b_j C_j}{1 + \sum_{k=1}^{1} b_i C_i}$$
(5)

where  $C_j$  is the equilibrium concentration of species j and  $C_i$  is the equilibrium concentration of all other adsorbing species in solution [18], while  $q_j^o$ ,  $b_j$  and  $b_i$  are constants obtained from singlecomponent studies for particular adsorbent–adsorbate system [11]. For two components, the model can be expressed as:

$$q_{e,1} = \frac{q_1^o b_1 C_{e,1}}{1 + b_1 C_{e,1} + b_2 C_{e,2}} \tag{6a}$$



**Fig. 3.** Extended Langmuir Model for RB21 in binary system. ( $\blacklozenge$ ) RB21+50 mg L<sup>-1</sup> RR195 experimental, ( $\blacksquare$ ) RB21+100 mg L<sup>-1</sup> RR195 experimental, ( $\blacktriangle$ ) RB21+150 mg L<sup>-1</sup> RR195 experimental; (-) RB21+50 mg L<sup>-1</sup> RR195 calculated, (---) RB21+100 mg L<sup>-1</sup> RR195 calculated, (---) RB21+150 mg L<sup>-1</sup> RR195 calculated.

$$q_{e,2} = \frac{q_2^0 b_2 C_{e,2}}{1 + b_1 C_{e,1} + b_2 C_{e,2}}$$
(6b)

where b and q are the Langmuir isotherm model parameters obtained suitably from Eq. (3) in the single solute system. Therefore, the amount of solute adsorbed were calculated Eqs. (6a) and (6b) and presented by Figs. 3 and 4 for each different system. The single-component isotherms for RB21 and RR195 have monolayer saturation capacities of 9.652 and 3.186 mg/g, respectively. It is obvious that RB21 have a higher absorbed on ZEC than RR195. because RB21 has a large size and heavy molecules, as shown in Scheme 1. Competition and interaction between these two reactive dyes would be significant and affect each component to a different extent. By comparing the experimental data for RR195 in binary solution 50, 100 and 150 mg L<sup>-1</sup> of RB21, it could be observed that the energy of adsorption between the specific absorbing species and the surface site is constant and equal for each site, there is no interaction between components and there is equal competition between species for adsorption sites, as seen in Fig. 4. This is clarified the basic assumption of Langmuir model, where no interaction and equal competition were occurred in the binary solution [19]. On the other hands, the extended Langmuir model could described the adsorption behaviour of RB21 in binary systems. The adsorption of RB21 was independent on the interaction of



**Fig. 4.** Extended Langmuir Model for RR195 in binary system. ( $\blacklozenge$ ) RR195+50 mg L<sup>-1</sup> RB21 experimental, ( $\blacksquare$ ) RR195+100 mg L<sup>-1</sup> RB21 experimental, ( $\blacktriangle$ ) RR195+150 mg L<sup>-1</sup> RB21 experimental; (-) RR195+50 mg L<sup>-1</sup> RB21 calculated, (---) RR195+100 mg L<sup>-1</sup> RB21 calculated, (---) RR195+150 mg L<sup>-1</sup> RB21

Table 4
Kinetics parameters for adsorption dyes from single and binary systems

Dye	Dose (mg L <sup>-1</sup> )	Pseudo-first-order $t = 0-45 \min$		Pseudo-second-order $t = 0-45$ min		
		$k_1 ({ m min}^{-1})$	R <sup>2</sup>	$k_2 (g m g^{-1} m i n^{-1})$	R <sup>2</sup>	
	50	0.120	0.9635	0.304	0.9999	
RB21	100	0.105	0.9573	0.039	0.9867	
	150	0.060	0.9476	0.025	0.9876	
	50+50	0.108	0.9854	0.052	0.9906	
RB21 + RR195	100+100	0.068	0.9724	0.010	0.9866	
	150+150	0.059	0.9257	0.003	0.9806	
	50	0.060	0.9546	0.196	0.9801	
RR195	100	0.049	0.9383	0.163	0.9941	
	150	0.035	0.7379	0.077	0.9860	
	50+50	0.059	0.9116	0.025	0.9801	
RR195 + RR21	100+100	0.058	0.9705	0.017	0.9800	
	150+150	0.042	0.9578	0.013	0.9892	



**Fig. 5.** Plots of concentration of RB21 versus time, using different solution concentrations on ZEC ( $\blacklozenge$ ) 50 mg L<sup>-1</sup>, ( $\blacksquare$ ) 100 mg L<sup>-1</sup>, ( $\blacktriangle$ ) 150 mg L<sup>-1</sup>.

RR195 in binary solution 50, 100 and 150 mg  $L^{-1}$  of RR195, as shown in Fig. 3.

## 3.4. Adsorption kinetics model for dyes onto ZEC

In this work, two kinetic models were employed to fit the experimental data. The first one was the pseudo-first-order kinetic model and the integral form of this model was expressed by the following equation

$$\ln(q_e - q_t) = \ln q_e - k_1 t \tag{7}$$

where  $q_e$  and  $q_t$  are the amounts of dye adsorbed on adsorbent at equilibrium and at any time *t*, respectively  $(mgg^{-1})$ ,  $k_1$  is the rate constant of pseudo-first-order model  $(min^{-1})$ ; and *t* is the contact time (min). The correlations of  $\ln(q_e - q_t)$  versus *t* in adsorption



**Fig. 6.** Plots of concentration of RR195 versus time, using different solution concentration on ZEC ( $\blacklozenge$ ) 50 mg L<sup>-1</sup>, ( $\blacksquare$ ) 100 mg L<sup>-1</sup>, ( $\blacktriangle$ ) 150 mg L<sup>-1</sup>.

from single and mixed dye solutions. In adsorption from mixed dye solutions, the total amounts of dyes adsorbed were used in data treatment [20]. The values of  $q_e$  were amounts of dyes adsorbed after a contact time of 90 min, as seen in Figs. 5 and 6. Adsorption data up to 45 min were used for kinetic calculations because the value of  $\ln(q_e - q_t)$  was sensitive to the error in  $q_t$  in the case of  $q_t \approx q_e$ .

The pseudo-second-order kinetic model is analyzed by Ho and McKay [21] based on the adsorption capacity and is expressed as follows:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \tag{8}$$

where  $k_2$  is the equilibrium rate constant of pseudo-second-order adsorption, mgg<sup>-1</sup>min<sup>-1</sup> and  $q_e$  is the equilibrium adsorption capacity, mgg<sup>-1</sup>. When the effect of initial concentration of both single (RB21, RR195) and binary solution (RB21with RR195, RR195 with RB21) were compared, it was observed that  $k_1$  and  $k_2$  values decrease with increasing initial concentrations. As shown in Table 4, the values of correlation coefficients of the rate constant of pseudo-first-order are less than  $R^2 < 0.980$ . As a result, the models which provided the best correlation with experimental data were the pseudo-second-order model.

# 4. Conclusions

In the present study the adsorption of reactive dyes, RB21 and RR195 on ZEC has been investigated in batch model. The equilibrium time of adsorption was found to be 90 min for ZEC. In the adsorption in single dye solutions, the adsorption isotherms of RB21 and RR195 onto ZEC were well described by the Langmuir adsorption equation. The single-component isotherms for RB21 and RR195 have monolayer saturation capacities of 9.652 and 3.186 mg/g, respectively. It is obvious that RB21 have a higher absorbed on ZEC than RR195. For single and binary solution of RR195, the basic assumption of Langmuir model could be accepted. The extended Langmuir model could described the adsorption behaviour of RB21 in binary systems. The adsorption kinetics for the two dyes from single dye systems, and the total amounts of dyes adsorbed from mixed dye systems, followed the pseudosecond-order kinetic equation. The adsorption rate constants were observed to decrease with increasing dosage of the reactive dves.

#### References

 P.C. Vandevivere, R. Bianchi, W. Verstaete, Treatment and reuse of wastewater from the textile wet-processing industry: review of emerging technologies, J. Chem. Technol. Biotechnol. 72 (1998) 289–302.

- [2] G. Annadurai, R.S. Juang, D.J. Lee, Use of cellulose-based wastes for adsorption of dyes from aqueous solutions, J. Hazard. Mater. B92 (2002) 263–274.
- [3] C. Hademal, F. Bocquillon, O. Zahraa, Decolourization of textile industry wastewater by the photocatalytic degradation process, Dyes Pigments 49 (2) (2001) 117–125.
- [4] A. Conneely, W.F. Smyth, G. McMullan, FEMS, Metabolism of the phthalocyanine textile dye remazol turquoiseblue by Phanerochaete chrysosporium, Microbiol. Lett. 179 (1999) 333–337.
- [5] D.T. Sponza, M. Işik, Decolorization and azo dye degradation by anaerobic/aerobic sequential process, Enzyme Microb. Technol. 31 (2002) 102–110.
- [6] E. Demirbas, M. Kobya, E. Senturk, T. Ozkan, Adsorption kinetics for the removal of chromium (VI) from aqueous solutions on the activated carbons prepared from agricultural wastes, Water SA 30 (2004) 533–539.
- [7] R.F. Moreira, N.C. Kuhen, M.G. Peruch, Adsorption of reactive dyes onto granular activated carbon, Latin Am. Appl. Res. 28 (1998) 37–41.
- [8] R.S. Juang, R.L. Tseng, F.C. Wu, S.H. Lee, Adsorption behaviour of reactive dyes from aqueous solutions on chitosan, J. Chem. Technol. Biotechnol. 70 (1997) 391–399.
- [9] E. Longhinotti, F. Pozza, L. Furlan, M.N.M. Sanchez, M. Klug, M.C.M. Laranjeria, V.T. Fávere, Adsorption of anionic dyes on the biopolymer chitin, J. Braz. Chem. Soc. 9 (5) (1998) 435–440.
- [10] B. Armağan, O. Özdemir, M. Turan, M.S. Çelik, The removal of reactive azo dyes by natural and modified zeolites, J. Chem. Technol. Biotechnol. 78 (2003) 725–732.
- [11] C. Valderrama, J.I. Barios, A. Farran, J.L. Cortina, Evaluating binary sorption of phenol/aniline from aqueous solutions onto granular activated carbon and hypercrosslinked polymeric resin (MN200), Water Air Soil Pollut. 210 (2010) 421–434.

- [12] F. Esenli, Quantitative analysis of heulandites-clinoptilolites in zeolite containing tuffs from Gordes region by X-ray diffraction, Publication of the Chamber of Geological Engineers of Turkey, 1993, pp. 42–49.
- [13] I. Arslan, I.A. Balcioglu, Degradation of commercial reactive dyestus by heterogenous and homogenous advanced oxidation processes: a comparative study, Dyes Pigments 43 (1999) 95–108.
- [14] M. Safarikova, I. Safarik, Magnetic solid-phase extraction of target analytes from large volumes of urine, Eur. Cell Mater. 3 (2) (2002) 192–195.
- [15] P. Peralta-Zamora, A. Kunz, N. Nagata, R.J. Poppi, Spectrophotometric determination of organic dye mixtures by using multivariate calibration, Talanta 47 (1998) 77–84.
- [16] V. Belesi, G. Romanos, N. Boukos, D. Lambropoulous, C. Trapalis, Removal of reactive red 195 from aqueous solutions by adsorption on the surface of TiO<sub>2</sub> nanoparticles, J. Hazard. Mater. 170 (2009) 836–844.
- [17] T. Sismanoglu, Removal of some fungicides from aqueous solution by the biopolymer chitin, Colloids Surf. A: Physicochem. Eng. Aspects 297 (2007) 38-45.
- [18] A.R. Khan, T.A. Al-Bahri, A. Al-Haddad, Adsorption of phenol based organic pollutants on activated carbon from multi-component dilute aqueous solutions, Water Res. 31 (8) (1997) 2102–2112.
- [19] K.K.H. Choy, J.F. Porter, G. McKay, Langmuir isotherm models applied to the multicomponent sorption of acid dyes from effluent onto activated carbon, J. Chem. Eng. Data 45 (2000) 575–584.
- [20] D. Shen, J. Fan, W. Zhou, B. Gao, Q. Yue, Q. Kang, Adsorption kinetics and isotherm of anionic dyes onto organo-bentonite from single and multisolute systems, J. Hazard. Mater. 172 (2009) 99–107.
- [21] Y.S. Ho, G. McKay, Pseudo-second order model for sorption processes, Process Biochem. 34 (5) (1999) 451–465.